

Tree mortality in a riparian forest at Rio Paraguai, Pantanal, Brazil, after an extreme flooding

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Received: January 07, 2004. Accepted: May 05, 2004

RESUMO – (Mortalidade de árvores em uma floresta ripária do rio Paraguai, Pantanal, Brasil, após uma cheia excepcional). Foi realizado um estudo numa mata ripária (rio Paraguai, Brasil) para verificar a mortalidade da vegetação após uma cheia excepcional, em 1995. Indivíduos com diâmetro a altura do peito ? 5cm foram marcados em 108 parcelas (10×10m) em 1994 e reamostrados em 1996. A taxa de mortalidade total foi de 4,1%. A mortalidade aumentou nas faixas topográficas mais altas, ao nível de comunidade, indicando que lugares onde as inundações são mais raras são mais afetados pelas grandes cheias.

Palavras-chave: Corumbá, inundação, perturbação, mata ripária, dique marginal

ABSTRACT – (Tree mortality in a riparian forest at Rio Paraguai, Pantanal, Brazil, after an extreme flooding). A study was conducted in a riparian forest (Rio Paraguai, Brazil) to verify the vegetation mortality after an exceptional flooding in 1995. Individuals with diameter at breast height ? 5 cm were sampled in 108 (10×10m) plots in 1994, and re-sampled in 1996. The total mortality rate was 4.1% per year. The mortality increased with the increasing of topographic positions, at community level, suggesting that places where the flooding is less frequent are more affected by extreme floods.

Key words: Corumbá, flooding, disturbing, riparian forest, riverbank

Introduction

Processes of recurrent natural disturbance are a very important component of the environmental setting of many plant communities. Regionally and within a landscape, the kinds of natural disturbance vary as function of the topography, substrate and the vegetation present (White 1979). In tropical environments, the maintenance of the diversity through attenuation of the effects of competitive exclusion is attributed to disturbance caused by tree fall. When the gaps arise in the process, local succession is re-conducted to the beginning (Connell 1978; Hubbell *et al.* 1999; Sheil & Burslem 2003).

In floodplains, inundations are an important disturbing factor. Besides the anoxia, it causes stream migration, erosion and deposition, creating new substrates and destroying old ones. The ability of species to colonize different environment, dispersion features, tolerance to inundation and shade tolerance

create a gradient that reflects the timing and frequency of floods (Junk 1989; White 1979). According to Naiman & Décamps (1997), the disturbing power and frequency of inundations are inversely proportional forces. Uncommon and exceptional flooding has more power to impact the riparian zones. According to Junk (1989), flooding tolerance can increase when a given tree becomes an adult. Even so, if the duration of the flooding is increased by an unusual inundation, the additional stress produced by anoxia could lead the less tolerant species to die. Nevertheless, there is little information about the effects of exceptional floods on riparian forests (Vervuren *et al.* 2003.)

The Rio Paraguai is the main river of the Brazilian Pantanal, and is subject to annual flooding. Exceptional floods are rare, with only three registered in the last century, in 1905, 1988 and 1995.

There is no information about patterns of mortality of trees in Pantanal riparian forests at these events. Studies on riparian forests of Rio Paraguai mainly deal

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with structure (Dubs 1994; Guarin-Neto *et al.* 1996; Veloso 1947) or the effects of inundation on the structure and distribution of species (e.g. Adámoli & Pott 1999; G.A. Damasceno-Junior *et al.* 1999; Nunes da Cunha & Junk 1999; Nunes da Cunha & Junk 2001; Prado *et al.* 1994).

This study is part of the masters research carried out by the first author between 1994 and 1996. The objectives were to describe the vegetation of riparian forests along the Rio Paraguai, around Corumbá, Mato Grosso do Sul State. During this period, an exceptional flooding occurred in 1995, providing the opportunity to study the mortality in these rare events, so, the aim of this study was to compare the mortality rates of the species in different topographic positions, after the exceptional flooding occurred in 1995, in a riparian forest of the Rio Paraguai. The mortality of individuals occurring in higher positions, where flooding is an unusual event, is expected to be higher than that of individuals occurring in lower positions submitted to annual flooding where the species are theoretically more tolerant to inundation and anoxia.

Material and methods

Study site - Located on the right riverbank of the Rio Paraguai downstream from the Rabicho ranch between 19°01'12" - 19°01'16"S and 57°27'01" - 57° 26'44"W in Pantanal of the Rio Paraguai subregion, municipality of Corumbá, Mato Grosso do Sul State, Brazil (Fig. 1).

The topography of the site is very flat, with a slope that reaches nearly 2cm.km⁻¹ in the north-south direction, the same direction that the Rio Paraguai flows. In this area, the river has numerous secondary channels, meanders, oxbow lakes, islands and a large area of seasonally flooded grasslands. The total inundation area of the Rio Paraguai has been varying from nearly 4,000 to 16,000km² between low and high water seasons during the last 95yr (Hamilton *et al.* 1996). The riparian forests are mainly located in convex parts of meanders, where sets of narrow riverbanks are deposited during the floods.

This area, like all of the Pantanal, has a distinct wet and dry season, a dry winter and rainy summer (Awa, according to Köppen system), with a mean annual precipitation of 1,070mm (Soriano 1997). At this area, the Rio Paraguai has an unimodal annual system of flooding. The flooding peak is reached in the dry season, with the lowest level occurring in the rainy months (Fig. 2). This occurs because the local inundation depends on the rains of the upper basin,

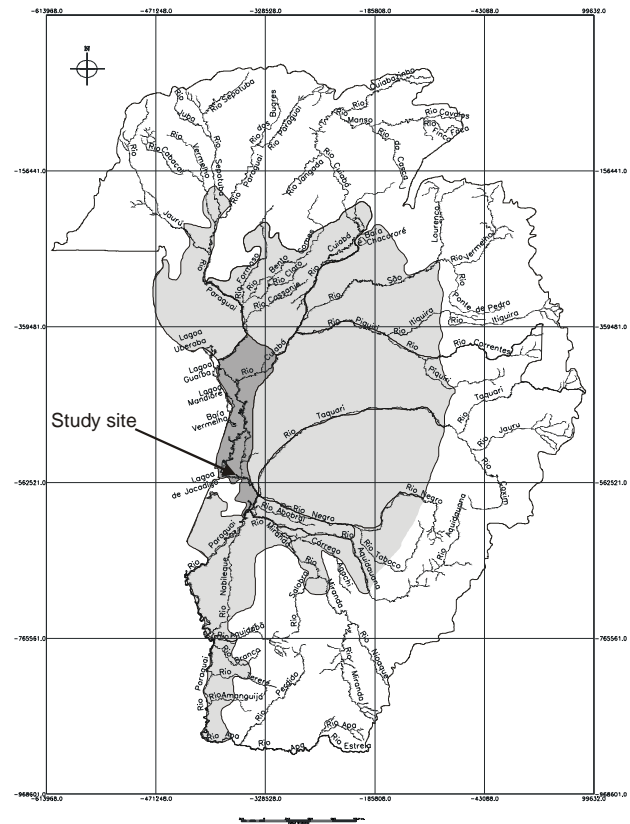


Figure 1. Location of the study area and Alto Paraguai basin, Brazilian Pantanal, according to the Pantanal subdivision made by Adamoli (1982) and modified by Silva & Abdon (1998). The darkest area is the Pantanal of the Paraguai sub-region.

whose peak flows take about three months to reach the study area. During normal flood periods, only higher riverbanks escape inundation, and in extreme floods, the entire floodplain is inundated.

In the Rio Paraguai floodplain, besides riparian forests, seasonally flooded grasslands and aquatic vegetation in the lagoons and water channels can be found (Bortolotto *et al.* 1999). The riparian forest varies from evergreen to alluvial semi-deciduous forest, where some deciduous species have a greater number of individuals.

Data collection - A grid of 108 contiguous 10×10m plots was systematically established to sample all locations of the riverbank. Because of the variation in the local micro topography, this grid covered all topographic locations of the chosen riverbank. All individuals with at least one branch of 5cm or more of DBH (diameter at 1.3m above the ground level) were sampled. The trees and shrubs were permanently labelled with a numbered aluminium tag. A voucher specimen of the different species was collected, dried and deposited in

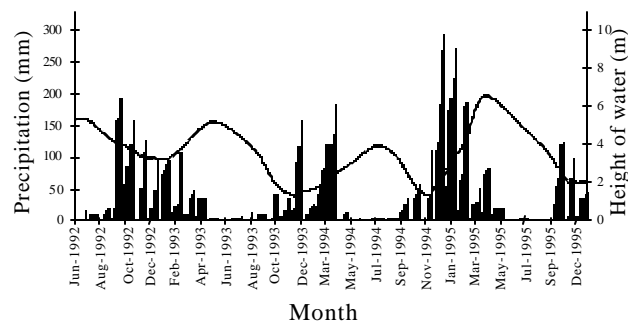


Figure 2. Daily levels of the Rio Paraguai on the Ladário gauge (line) and monthly-accumulated precipitation (bars) in Corumbá between June/1992 and December/1995. Data provided by the Brazilian Navy and Air Force, respectively.

the Herbarium of the Universidade Federal de Mato Grosso do Sul (COR), with duplicates to the Universidade Estadual de Campinas (UEC) and Embrapa (CPAP).

The highest water mark, left by the 1995 flooding on the trunk of each tree, was measured. A staff gauge in the study site recorded the inundation level variation during three months. This variation was the same as shown by the Ladário gauge used by the Brazilian navy which has been collecting daily data since 1900 ($R^2 = 99.85\%$, $P < 0.001$). Because of this, and to facilitate the historical analysis of flooding on each individual, the measurements of water mark height were transformed into topographic localization of each tree in reference to the zero level of the Ladário gauge. Since the object of this analysis was to investigate the

mortality in the different topographic positions, and the original plots were often quite heterogeneous with respect to flooding level, these plots were disregarded. To perform the analysis, the sampling units used were bands defined by their height above the zero level of the Ladário gauge. Each 40cm height interval above the zero mark was considered to be a unit.

The period between 1974 and 1996 was considered for an analysis of how much time each topographic position of the surveyed riverbank remained flooded. This was chosen because, after 1974, the inundation regimen of the Rio Paraguai experienced a change, and is now in a period of more regular flooding after a long drought in the 1960's (Fig. 3).

All individuals were measured in 1994, and re-sampled in 1996. To verify if a given individual had died, part of its bark was removed to check if there was humidity in the tissues inside. This was performed for fallen as well as standing trees.

Annual mortality rate was calculated using the model described in Sheil *et al.* (1995) and Sheil & May (1996), where: $m = 1 - (N_t/N_0)^{1/t}$ (N_0 is the number of trees at the first measurement and N_t is the number of survival trees at the second measurement, and t is the time between the measurements, 2 yr in this study) for each species and $m_{all} = \frac{1}{n} \sum_{i=1}^n [n_{i0}(1 - m_i)]^{1/t} / n_{i0}$ is for all species (where n_0 is the initial number of individuals and m_i is the mortality rate of the i^{th} species). Both results were multiplied by 100 and presented in mean percentage per yr.

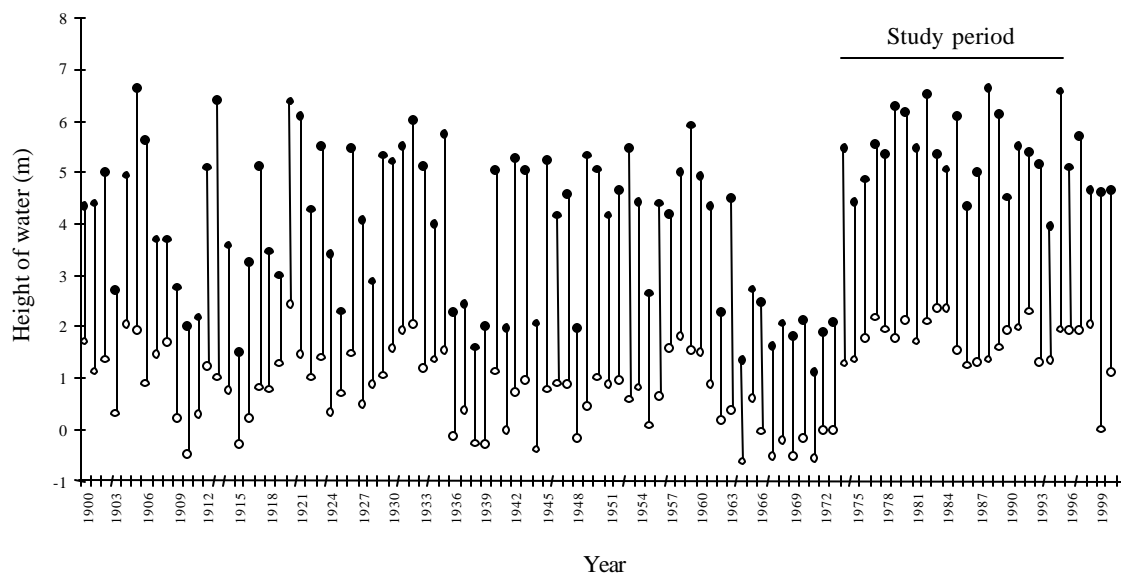


Figure 3. Annual maximum (full circle) and minimum (empty circle) levels of the Rio Paraguai on the Ladario gauge during the entire 20th century, showing the period chosen for this survey, Mato Grosso do Sul, Brazil. Zero is at 83m above sea level. Data from the Brazilian Navy.

To verify the differences between the numbers of dead individuals in the different diameter classes for each species, a chi-squared test was applied. The proportion of dead individuals in the lower diameter class was compared with the proportion of dead individuals in the higher class. For these comparisons, only the initial measures were considered; no recruitment analysis was performed. A regression analysis was applied to verify if the mortality increased with increasing topographic bands (Zar 1996). In this analysis, the mortality index was calculated for all individuals in each topographic band considered.

Results

The areas up to 3.8m high in the riverbank were inundated in all 23yr considered, being flooded from 120 to 250 days per yr (Fig. 4). The areas between 4.2 and 4.6m high were flooded from 19 to 22yr in the 23yr analyzed, being from 10 to 20 consecutive years and remained inundated about 90 to 180 days in flooded years. Areas 5 to 6m were occasionally inundated, with at least two consecutive yr without inundation, and a maximum of 9yr with consecutive flooding, remaining 11 to 90 days with water, when inundated.

A total of 695 individuals distributed into 37 species were tagged in 1994 (Tab. 1). The mortality rate was 4.11% per yr, corresponding to 56 dead individuals in 2yr. Among the individuals with 5-15cm DBH stems, the mortality was 4.6% per yr, and among the stems over 15cm DBH, it was 3.3% per yr (Tab. 2).

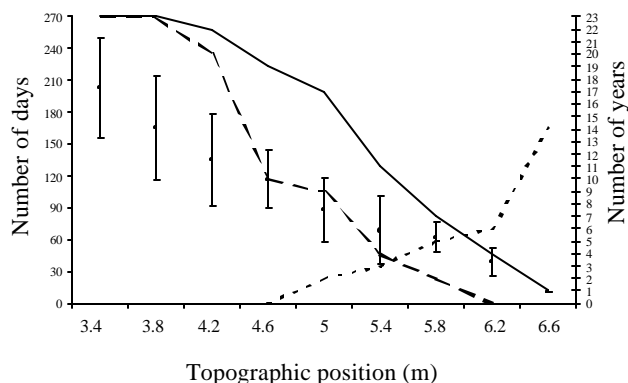


Figure 4. Topographic positions in the riparian forest of Rio Paraguai, Corumbá, Mato Grosso do Sul State, Brazil, between 1974 and 1996 showing: (—) total number of years inundated; (---) number maximum of consecutive years inundated; (I) mean and standard deviation of days y^{-1} flooded (only for flooded years); (---) maximum number of consecutive years without flooding.

Table 1. Species sampled in the riparian forest of the Rio Paraguai in Corumbá, Mato Grosso do Sul State, Brazil, with respective botanical families, growth habits and number of individuals (N).

Family/Scientific name	Habitat	N
ARECACEAE		
<i>Acrocomia aculeata</i> (Jacq.) Lodd.	Palm	3
<i>Bactris glaucescens</i> Drude	Palm	6
BIGNONIACEAE		
<i>Tabebuia heptaphylla</i> (Vell.) Tol.	Tree	7
CAPPARACEAE		
<i>Crataeva tapia</i> L.	Tree	50
CECROPIACEAE		
<i>Cecropia pachystachya</i> Tréc	Tree	31
CHRYSOBALANACEAE		
<i>Couepia uiti</i> (Mart. & Zucc.) Bth.	Shrub	7
EUPHORBIACEAE		
<i>Alchornea castaneifolia</i> (Willd.) A. Juss.	Shrub	12
<i>Alchornea discolor</i> Poepp.	Tree	10
<i>Sapium obovatum</i> Kl.	Tree	2
FLACOURTIACEAE		
<i>Casearia aculeata</i> Jacq.	Shrub	10
HIPPOCRATEACEAE		
<i>Salacia elliptica</i> (Mart.) Peyr.	Shrub	1
LAURACEAE		
<i>Ocotea diospyrifolia</i> (Meisn.) Mez	Tree	56
LEGUMINOSAE - CAESALPINIOIDEAE		
<i>Hymenaea courbaril</i> L. var. <i>Stilbocarpa</i> (Hayne) Y.T. Lee & Langenh.	Tree	1
LEGUMINOSAE - MIMOSOIDEAE		
<i>Albizia inundata</i> (Mart.) Barneby & J.W. Grimes	Tree	19
<i>Inga vera</i> Willd. ssp. <i>affinis</i> (DC.) Pennington	Tree	226
<i>Zygia inaequalis</i> (H.B.K.) Pitt.	Shrub	5
LEGUMINOSAE - PAPILIONOIDEAE		
<i>Andira inermis</i> H.B.K.	Tree	7
<i>Bergeronia sericea</i> Mich.	Shrub	2
<i>Pterocarpus micheli</i> Brit.	Tree	10
MELIACEAE		
<i>Trichilia stellato-tomentosa</i> Kze.	Shrub	4
MELASTOMATACEAE		
<i>Mouriri guianensis</i> Aubl.	Shrub	2
MORACEAE		
<i>Brosimum gaudichaudii</i> Tréc.	Shrub	1
MYRTACEAE		
<i>Eugenia egensis</i> DC.	Shrub	8
<i>Eugenia cf. polystachya</i> Rich.	Tree	21
<i>Myrcia cf. mollis</i> (H.B.K.) DC.	Shrub	21
NYCTAGINACEAE		
<i>Neea hermaphrodita</i> S. Moore	Shrub	13
POLYGONACEAE		
<i>Coccoloba cf. alagoensis</i> Wedd.	Tree	12
<i>Ruprechtia brachysepala</i> Meisn.	Tree	9
<i>Triplaris gardneriana</i> Wedd.	Tree	70
RUBIACEAE		
<i>Genipa americana</i> L.	Tree	1
<i>Tocoyena formosa</i> C. & S.	Shrub	4

continue

Table 1 (continuation)

Family/Scientific name	Habitat	N
SAPINDACEAE		
<i>Cupania castaneaefolia</i> Mart.	Tree	1
SAPOTACEAE		
<i>Pouteria glomerata</i> (Miq.) Radlk.	Tree	10
STERCULIACEAE		
<i>Guazuma ulmifolia</i> Lam.	Tree	7
ULMACEAE		
<i>Celtis pubescens</i> (H.B.K.) Spreng.	Shrub	8
VERBENACEAE		
<i>Vitex cymosa</i> Bert.	Tree	7
VOCHYSIACEAE		
<i>Vochysia divergens</i> Pohl	Tree	31

The mortality rates in each topographic region increased from low to high positions, indicating that after the 1995 extreme flood, the trees that were at higher topographic positions had a proportionally greater number of dead individuals (Fig. 5). None of the species with four or more dead individuals showed differences in mortality at different topographic position. Nevertheless, except for *Pterocarpus michelli* and *Eugenia egensis*, the other species with one dead individual were found predominantly in high bands where there were at least two consecutive years without flooding in the 23yr analysed (Tab. 3). That is

why the general mortality was proportionally greater in the higher sites. No mortality was recorded for individuals occupying the sites below 3.8m (Fig. 5).

Among the species with more than 20 individuals, the mortality rate varied from no mortality in *Crataeva tapia*, to 32.8% per yr in *Cecropia pachystachya*. For *Inga vera* ssp. *affinis*, the most abundant species, mortality was 3.15% per yr. *Triplaris gardneriana*, *Ocotea diospyrifolia* and *Vochysia divergens* showed mortality rates similar to *Inga vera*. There were no

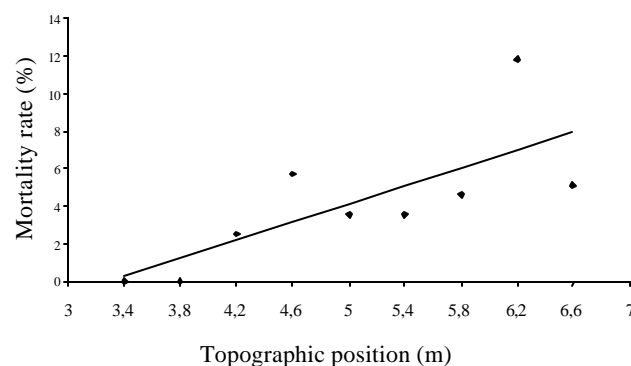


Figure 5. Mortality rate per year at each topographic position in a riparian forest of the Rio Paraguai, Corumbá, Mato Grosso do Sul State, Brazil, between the years 1994 and 1996. The points are the rates obtained and the straight line are from regression analysis ($R\text{-Sq} = 55.8\%$ and $P = 0.02$). The total number of individuals at each band was 11, 7, 100, 134, 158, 157, 100, 18 and 10, respectively.

Table 2. Mortality of species in the riparian forest of the Rio Paraguai, Corumbá, Mato Grosso do Sul State, Brazil, between 1994 and 1996. T - total number of individuals; T 5-15 and T >15 - total of individuals with DBH between 5 and 15cm, and greater than 15cm; M - number of dead individuals after two years; M 5-15 and M >15 - number of dead individuals with DBH between 5 and 15cm, and greater than 15 cm; m (% y^{-1}) total mortality rate; m5-15 (% y^{-1}) - mortality rate among the individuals with DBH between 5 and 15cm; m >15 (% y^{-1}) - mortality rate among individuals with DBH greater than 15cm. Only the species which had dead individuals are presented in detailed way.

Species	T	M	T 5-15	M 5-15	T >15	M >15	m (% y^{-1})	m (% y^{-1}) 5-15	m (% y^{-1}) >15
<i>Inga vera</i> ssp. <i>affinis</i>	226	14	55	6	171	8	3.15	5.61	2.37
<i>Triplaris gardneriana</i>	70	4	60	3	10	1	2.90	2.53	5.13
<i>Ocotea diospyrifolia</i>	56	4	42	3	14	1	3.64	3.64	3.64
<i>Cecropia pachystachya</i>	31	17	20	11	11	6	32.80	32.92	32.58
<i>Vochysia divergens</i>	31	2	18	1	13	1	3.28	2.82	3.92
<i>Eugenia</i> cf. <i>polystachya</i>	21	9	21	9	0	0	24.41	24.41	-
<i>Neea hermaphrodita</i>	13	1	13	1	0	0	3.92	3.92	-
<i>Pterocarpus michelli</i>	10	1	8	1	2	0	5.13	6.46	0.00
<i>Celtis pubescens</i>	8	1	8	1	0	0	6.46	6.46	-
<i>Eugenia egensis</i>	8	1	8	1	0	0	6.46	6.46	-
<i>Trichilia stelato-tomentosa</i>	4	1	4	1	0	0	13.40	13.40	-
<i>Sapium obovatum</i>	2	1	2	1	0	0	29.29	29.29	-
Other species	215	0	174	0	41	0	0.00	0.00	0.00
Total	695	56	433	39	262	17	4.11	4.61	3.30

Table 3. Mortality rates (% y^{-1}) for all species with dead individuals in each topographic position considered in a riparian forest of the Rio Paraguai Corumbá, Mato Grosso do Sul State, Brazil, between 1994 and 1996. Species indicated with “-” didn’t occur in the band.

Species	Topographic position (m)								
	3.4	3.8	4.2	4.6	5	5.4	5.8	6.2	6.6
<i>Inga vera</i> ssp. <i>affinis</i>	0.0	0.0	4.7	0.0	4.0	2.7	3.9	0.0	13.4
<i>Triplaris gardneriana</i>	0.0	-	1.9	4.4	3.9	0.0	0.0	-	-
<i>Ocotea diospyrifolia</i>	0.0	-	10.6	8.7	0.0	0.0	0.0	0.0	-
<i>Eugenia</i> cf. <i>polystachya</i>	-	-	10.6	34.5	22.5	29.3	29.3	-	-
<i>Eugenia egensis</i>	-	-	100.0	-	0.0	0.0	0.0	-	-
<i>Pterocarpus micheli</i>	-	-	0.0	0.0	0.0	100.0	0.0	-	-
<i>Cecropia pachystachya</i>	-	-	-	46.5	13.4	33.3	34.5	29.3	-
<i>Neea hermaphrodita</i>	-	-	-	0.0	0.0	-	8.7	-	-
<i>Vochysia divergens</i>	-	-	-	0.0	10.6	0.0	10.6	0.0	-
<i>Sapium obovatum</i>	-	-	-	100.0	-	0.0	-	-	-
<i>Celtis pubescens</i>	-	-	-	-	-	-	0.0	18.4	0.0
<i>Trichilia stelato-tomentosa</i>	-	-	-	-	-	-	0.0	29.3	-

differences in mortality when trees with 5-15cm DBH stems were compared with those with DBH greater than 15cm for any species (chi-square test $p > 0.08$ - Tab. 2). In other words, the flooding affected the trees equally regardless of their age.

Cecropia pachystachya showed the highest annual mortality in the years studied. *Eugenia* cf. *polystachya* showed the second highest mortality rate among the species with 20 or more individuals (24.4% per yr). All the individuals of *Eugenia* cf. *polystachya*, which is an understory species, measured less than 15cm DBH. The remaining species reported in Tab. 1 with low number of total individuals or of dead individuals were cited only for calculation of the total mortality index at the community level.

Discussion

There are few studies dealing with mortality in tropical environments, especially in riparian zones. Mortality rates for tropical trees cited in the literature reach around 2% per yr, being higher under stress caused by natural (like El Niño) and human impacts, reaching 6 to 7% per yr or even 9% per yr in edges of fragmented tropical forests (Aiba & Kitayama 2002; Burslem *et al.* 2000; Laurance *et al.* 1998; 2001; Mesquita *et al.* 1999). Using the mortality index of Swaine (1987), which has little difference in terms of absolute results from the one used here, Felfili (1995) found 3.5% in 6yr of observation in a gallery forest in central Brazil; in three seasonally flooded forests with different ages, Campbell *et al.* (1992) found 1.99%, 1.56% and 1.7% during a normal year of observation.

The annual mortality rate obtained for trees in the Rio Paraguai was high (4.11% per yr) compared with other tropical areas, as cited above.

The real annual mortality in the year 1995 could have in fact been higher than presented here. This is because the index obtained is from a mean of two years (1994 and 1995), aggregating the mortality in 1994 with the mortality in 1995, when the exceptional flood occurred. In spite of being a mean from two years, this high rate may be associated with the exceptional flooding level reached by the Rio Paraguai in 1995, which was the second highest since 1974. Flooding is a regular event in the study area, and hence, theoretically, its effects are absorbed by the community. Nevertheless, in exceptionally high floods, some phenomena work in a different way than in lower ones. The water speed is faster; the sedimentation rate could be high, reaching up to 10cm of sand in some points (Damasceno-Junior pers. obs.); edge points of the riverbank lose some of their soils, and many trees can fall as a consequence; and the time during which the highest topographic parts remain inundated is longer. Certainly, these aspects contributed to the high mean mortality obtained, especially on the high parts, where the effects of stream velocity and water permanence were larger.

Differences between mortality in low and high classes of DBH have been found in situations of stress caused by dry period (Aiba & Kitayama 2002; Condit *et al.* 1995) and with different histories of past disturbance or edge effect (Oliveira-Filho *et al.* 1997). For the exceptional flooding studied, there is no tendency for high mortality occurs among young

individuals. This is probably because the event affects the entire populations of trees with the same intensity.

The mortality was high; especially for *Cecropia pachystachya* and *Eugenia cf. polystachya*. In *C. pachystachya*, the values may be easily associated with the exceptional flooding in 1995. *C. pachystachya* is a pioneer species with a weak trunk structure. Although the *causa mortis* was not recorded, many individuals of this species fell during the inundation due to the water current. Condit *et al.* (1995) found high mortality for *Cecropia obtusifolia* during an unfavourable dry period, and Felfili (1995) found the highest value of mortality for pioneer species in a gallery forest. For *Eugenia cf. polystachya*, the high mortality can be the result of its understorey location, which is vulnerable to falling branches or whole trees, the fast water current, and the floating mats carried by the water. In tropical rain forests, understorey species also had high mortality rates (Condit *et al.* 1995; Swaine *et al.* 1987).

Species like *Triplaris gardneriana*, *Vochysia divergens* and *Ocotea diospyrifolia* did not seem to have been affected by the exceptional high flood. *Triplaris gadneriana* occurs predominantly in low places and the other two are indifferent to topographical position (Damasceno-Junior pers. obs.). Besides, there were no dead trees from any species below 3.8m sites, that were inundated all the 23yr considered. Probably they are well adapted to being flooded during most of the months every year (see Fig. 4).

Although for some species like *Cecropia pachystachya* and *Eugenia cf. polystachya*, the exceptional flood had more impact, there were no differences between mortality in lower and higher locations for any species. This high flood increased the flooding time and the effects of the speed of the water flow across the whole occurrence band for each species. These phenomena increased mortality throughout the whole topographic band of occurrence in the species affected, and not in the bands artificially defined for this study. Because of this, the tendency to have increasing mortality in the higher topographic positions is only detectable at the community level and not for any isolated species. This is a very subtle effect, in which the effects of the increase in flooding time and speed of water flow become progressively stronger with increasing height of the location.

The higher mortality observed in higher locations may be, in part, due to the possibility that species located at the same height have similar tolerance of

flooding. Nevertheless, the events were apparently not sufficiently strong to cause intense reductions in all populations, since in the majority of species from higher locations, a small number of individuals died. This may be to the fact that the individuals studied were adults, which according Junk (1989), would be more resistant to flooding.

In conclusion, expectations regarding increasing mortality in higher topographic locations were confirmed mainly at the community level, but were not detectable at the species level. The exceptional flood affected the system because of its unusual occurrence, especially in the parts of the community where the flood does not reach every year. The flood increases the chances of mortality for more species in higher locations, but not necessarily in great numbers.

Another point is that this flood also modified the configuration of the landscape. Vervuren *et al.* (2003) suggest that extreme floods can determine the distribution of species for many years on the Rhine River. In this study the data do not permit take conclusions about this subject, but it should be investigated in future studies to complement the data presented here. This will point to new directions to the study of the role played by inundations on the dynamics of seasonally flooded vegetation in the Pantanal.

Acknowledgements

We are grateful to the Brazilian Navy for the permission to work on the Rabicho farm; to Drs. Arnildo Pott and Ronald Fortney, for the first revision of this manuscript; to Antonio Glauter Cavalheiro, Denilson A. dos Santos and the students Diogenes P. Rodriguez, Luiz Claudio Baracat, Ligia Teixeira, and Andrea Sanches who helped during the field work; two anonymous reviewers; CAPES (Coordenação de Aperfeiçoamento de Pessoal de Nível Superior) and Universidade Federal de Mato Grosso do Sul who sponsored this work.

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